BAEOXYRON ALPINUM (L.) T.V. EGOROVA (CYPERCAEAE) IN THE POLISH LOWLANDS: DISTRIBUTION, POPULATION DECREASE AND IMPLICATIONS FOR CONSERVATION

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ABSTRACT
A total of 47 localities of Baeothryon alpinum, hosting population of more than 100 000 shoots, were recorded in the lowland part of Poland during field surveys in the years 2003-2009. Among them were 25 populations discovered for the first time. Out of 57 sites of the species known from literature and unpublished (including herbarium) sources, 35 were not confirmed during the survey, 27 of them being definitely extinct. B. alpinum shows a clear pattern of distribution in Poland, with three main areas of occurrence: 1) the north-easternmost Poland (Lithuanian Lake District with the adjacent parts of the Masurian Lake District and the upper Biebrza river valley in North Podlasie Lowlands), which is part of the species boreal main range; 2) scattered localities in north-western Poland; 3) mountain mires at higher altitudes in the Sudetes and Tatra mountains and the adjacent part of southern Poland. The main aggregation of localities is found in Augustów Forest (including the Sejny Lakeland and Wigry National Park), and in the Góry Sudeckie region with adjacent areas. The biggest Polish population in the “Kobyła Biel” fen near Augustów consisted of several dozens of thousands of shoots. The Lithuanian Lake District is an area of general importance for the conservation of B. alpinum in Poland. The species is threatened, first of all, due to secondary succession (mires overgrowing with shrubs, trees and reed) and requires conservation measures as well as establishing nature reserves in places where it occurs. The degree that B. alpinum decreases in number is strikingly different in particular regions of Poland – it has lost most of its localities in north-western Poland and in Masurian Lake District, while in the Lithuanian Lake District and the upper Biebrza valley there are minor losses only. Depending on the region (from the west to the east and from the south-west to the north-east), the species should be given extinct or critically endangered (regions of north-western and southern Poland), endangered (Masurian Lake District), vulnerable (North Podlasie Lowlands) and near threatened (Lithuanian lake District) status. Although the disappearance of the populations beyond the species main range is a common phenomenon, the presented pattern is man-related and connected with differences in land management.

KEY WORDS: Baeothryon alpinum, Trichophorum alpinum, Cyperaceae, threatened species, distribution, conservation status.

INTRODUCTION

Baeothryon alpinum (L.) T.V. Egorova (= Trichophorum alpinum (L.) Pers., Scirpus hudsonianus (Michx) Fernald, Eríophorum alpinum L.) is a perennial mire plant forming loose mats. Shoots 10-40 cm, terminate with a single spikelet, 5-7 mm long and ca. 3 mm wide, ellipsoid to lanceolate. Perianth bristles 4-6, cotton-like in appearance, white, smooth, up to 25 mm long (DeFilipps 1980). This circumpolar species is bound primarily to the boreal (taiga) zone and the mountain areas southwards. In Europe it is widespread in Scandinavia, northern Russia and the Baltic countries. Its main distribution range extends to Denmark and north-easternmost Poland. Scattered lowland localities are known from northern Germany, western part of Poland and the upper Don area in central Russia, whereas quite numerous mountain populations are found in mountain ranges of central and southern Europe, e.g. the Alps, Pyrenees, Carpathians, Sudetes and Apennines (Hultén and Fries 1986a, b).

The distribution of Baeothryon alpinum in Poland shows a distinct pattern, with three separate areas of occurrence. The main area, which is considered as part of the species boreal range, covers the north-easternmost part of the Polish Lowlands (Lithuanian Lake District with the adjacent parts of the Masurian Lake District and the upper Biebrza river valley). The remaining two areas of occurrence are the post-glacial landscape of north-western Poland and the mountain areas of southern Poland (the Sudetes and Tatra Mountains) with the adjacent, southernmost part of Polish
lowland (Żukowski 1969; Zając and Zając 2001). The reason
for the observed gap in the species distribution in north-
ern Poland (spreading between the Parsęta and Gwda ri-
ers in Pomerania and Śniardwy Lake in Masuria) was a
subject of interest of former botanists. Czubiński (1950)
presumed that originally the species had a constant distri-
bution in northern Poland, but the area has been divided by
several glacier encroachments at the end of the last ice age.

In Poland, the species grows in various mire types, in-
cluding Sphagnum-dominated poor fens, brown moss-small
sedge rich fens (e.g. spring fens) and even in pine-birch fen
woodlands (Pańczyński 1975; Sokołowski 1988, 1996; Ty-
szkowski 1993; Dajdok and Kački 1999; Łachacz and Ole-
siński 2000a; Pawliński 2008a). During the last century
the localities of *B. alpinum* scattered in Europe beyond the
species main range have been vanishing as a result of drain-
age, secondary succession and management intensification
(Ditě and Hoskovec 2005). In Poland, the species has
lost most of its localities in the western part of the country
and is considered vulnerable (VU category – Żukowski
2001). In the countries adjacent to Poland to the south
and west, the species is considered critically endangered
(Slovačka – Ditě and Hoskovec 2005) or endangered (the
provinces Mecklenburg-Vorpommern and Schleswig-Holste-
in in north-eastern Germany – Ingelöf et al. 1993; Czech
Republic – Ditě and Hoskovec 2005). In all of the publi-
shed Polish regional lists the species is listed as endan-
gered (Western Pomerania – Żukowski and Jackowiak 1995;
Polish Carpathians – Piękos-Mirkowa and Walusiak 2008),
critically endangered (Opole Province – Dajdok and Kački
2002) or extinct (Wielkopolska region – Jackowiak et al.
2007). North-eastern Poland is considered the most impor-
tant region for the species in Poland (Żukowski 1969,
2001). Nevertheless, there is no clear evidence of that and
the population size is unknown. Moreover, the status of
about a dozen of localities recorded before the Second
World War in the Masurian Lake District (former East
Prussia) remains unknown. In the whole country, there is
little confirmed information on the disappearance of par-
cular localities. It pertains mainly to populations existing
in western Poland (Żukowski 2001) and in the mountains
(e.g. near Morskie Oko – Piękos-Mirkowa and Walusiak
2008). Only the population of *B. alpinum* in north-eastern
Poland whose extinction is sufficiently documented existed in
“Sypkowo” nature reserve near Giżycko (Polakowski 1959; Endler and Dziedzic 1991).

The aim of this study is to determine the distribution dy-
namics of *Baeothryon alpinum* in Poland and assign the
conservation status in particular regions and in the whole
country. In order to undertake conservation measures, re-
liable data on the species distribution dynamics, population
size, threats and the status of localities are needed.

**METHODS**

A review of literature and all the Polish herbaria collec-
tions was carried out to reveal the historical data on the
species occurrence in the lowland part of Poland (exclu-
ding mountain areas). All the historical localities in north-
ern Poland were surveyed in the years 2002-2009 to
determine whether the species still occurs there or not. Mo-

nover, suitable habitats were checked in order to find unk-
known populations. In all the localities confirmed or dis-
covered, the abundance of the species was estimated and the
threats were listed. In the case of extinct or probably
extinct localities, the possible reasons for extinction were
determined. Herbarium specimens collected during the
field survey were deposited in the Herbarium of Faculty of
Biology (WA) at University of Warsaw.

Every locality was assigned an appropriate ATPOL grid
square (Zając 1978; Zając and Zając 2001) and rated to ap-
propriate geographical region (Kondracki 2001). When po-
ssible, apart from the 10x10 km squares, smaller 5x5 km

quarters (one-fourth of the big square) were applied. For
example, letter “A” denotes north-west quarter, letter “B”
denotes north-east quarter and so on. In cases when several
local populations existed within one large peatland, they
were treated as one locality (unless they were located in
different ATPOL squares).

As a result of the field, herbarium and literature survey
of *Baeothryon alpinum* distribution in the Polish Lowlands,
a map of distribution was prepared. Mountain localities
were also included, based on the map of Zając and Zając
(2001, and modified using the data presented by Piękoś-
Mirkowa and Walusiak 2008).

**RESULTS**

ATPOL grid square codes are given next to the locality
name. Explanations of abbreviations: BPN – Biebrza Na-
tional Park; comm. – commune; Ex – sites extinct due to
habitat destruction; Ex? – sites most probably extinct; pers.
comm. – personal communication; res. – nature reserve;
sett. – settlement; v. – village; WPN – Wigry National
Park; * – localities threatened due to secondary succession
(trees, shrubs or reed); ** – localities severely threatened
due to secondary succession; ! – Paweł Pawliński unpub-
lished data. Abbreviations of the names of herbaria follow
Mirek et. al. (1997). Population size is presented using in-
tervals: 1 – several up to several dozens of shoots, 2 – se-
veral hundreds of shoots, 3 – several thousands of shoots; 4
– more than 10 000 shoots. In the case of localities recor-
ded before the Second World War by German botanists,
the former names are cited in [brackets]. Dates of the last
confirmed field sightings are given using **bold**.

**List of reliable localities in Poland excluding mountain areas**

**SZCZECIN COAST**

**AB95B** meadows [Maduewiesen] near Miedwiecko settl.
next to Zielieniewo v. and Miedwie Lake, Kobylanka or
Stargard Szczeciński comm. (Holzfuß 1940; Czubiński
1950; Żukowski 1969). Ex (Żukowski 2001). **1850**

**WEST POMERANIA LAKE DISTRICT**

**BB32** near Gładzino v., Rąbino comm. (Czubiński 1950;
Żukowski 1969). Ex (Żukowski 2001). **unknown (befo-
re ca. 1939)**

**BB43** Połczyn-Zdrój town, Połczyn-Zdrój comm. (Czubi-

**SOUTH POMERANIA LAKE DISTRICT**

**BC05D** “Trzęsacz” [Nakeler Faulen Bruch] fen in “Wielki
Bytyń” res., east of Wielki Bytyń Lake [Grosser Böthin-
see, Walcz comm. (Abromeit 1928; Frase 1930; Abromeit et al. 1931-40; Czubiński 1950; Balcerkiewicz and Markowski 1969; Źukowski 2001; W. Źukowski, pers. comm. in 2008). Ex? (W. Stachnowicz, pers. comm. in 2009) due to overgrowing with trees and reed. 1983

**LUBUSZ LAKE DISTRICT**

**SILESIAN LOWLANDS**

**SILESIAN HIGHLANDS**
**DF22C** Sucha Góra, district of Bytom city, Bytom comm (Zapałowicz 1906; Źukowski 1969). Ex (Źukowski 2001). 1875

**MASURIAN LAKE DISTRICT**
**EB68B** Masurian Scenic Park, “Bagno Mulaste” (Bagno Muliste) fen bordering small lakes south of Zgon [Sgonn] v., Piecki or Świętańo comm. (Abromeit 1928; Abromeit et al. 1931-40, Źukowski 1969). Ex! († 2009), but small patches of suitable habitat still exist. 1922

**FA83D** peatland near Mażucie [Masutschen] v., Goldap comm. (Abromeit 1928; Abromeit et al. 1931-40). Ex! († 2009) due to peat extraction, drainage and overgrowing with trees. 1925 or 1926

**FA83B** peatland near Wilkajcie [Wilkatschen] v., Goldap comm. (Abromeit et al. 1931-40). The paper containing the original data of F. Brauer (see Abromeit et al. 1931-40) has not been revealed in the present study. Ex! († 2009) due to peat extraction, drainage and overgrowing with trees. 1931

**FA83C** peatland near Audyniszki [Audinischken] v., Banie Mazurskie comm. (Abromeit et al. 1931-40). Ex! († 2009) due to peat extraction, drainage and overgrowing with trees and shrubs. This locality is not mentioned in the original paper of Abromeit (1928), which includes other nearby recordings of the species by R. Bühle (see Abromeit et al. 1931-40). 1925

**FA91C** fen in the place of non-existing lake near Kalskie Nowiny [Kehlerwald] v. between Węgorzewo and Stęgiel, Węgorzewo comm. (Führer 1928; Abromeit 1928; Abromeit et al. 1931-40; Źukowski 1969). Ex! († 2009) due to drainage and overgrowing with trees. between 1917 and 1920

**FA94A** peatland [Johannisberger Moor] south-west of Janowo [Johannisberg] v., Goldap comm. (Kauhohen and Range 1906). Ex! († 2009) due to overgrowing with trees and peat extraction. 1905

**FA94B** poor fen bordering a small lake near the top of the Tatarska Góra [Tataren Berg, Friedrichower Berg] hill, Goldap comm. (Schultz 1892; Abromeit 1898; Kauhohen and Range 1906, Gross 1912; Abromeit et al. 1931-40). Ex (Lachacz and Oleśiński 2000b, ! 2002-2004, 2006, 2009.), most probably due to acidification (but small patches of suitable habitat still exist). Gross (1910) erroneously cited the name of the hill as “Seesker Berg” (see Lachacz and Oleśiński 2000b). 1908 or 1909

**FB11B** fen in “Spytkowo” [Spieckergen] res., Giżycko comm. (Gross 1910; Abromeit et al. 1931-40, Polakowski 1959). Ex (Endler and Dziedzic 1991) due to drainage and overgrowing with trees and shrubs. ca. 1957

**FB26D** fen bordering a small lake [Wisnzy-See, Wisnzysee] that does not currently exist, in a forest north-east of Kleczewo [Kleszienow] v. and east of Nory settl., Wieliczki comm. (Grütter 1987; Abromeit et al. 1931-40; Źukowski 1969, ! 2009: 3). 2009**

**FB26D** fen in a forest west of Dorsze [Dorschen] v., Wieliczki comm. (Abromeit et al. 1931-40). Ex! († 2009) due to drainage, forest development and mire acidification. The paper containing the original data of G. Rehse (see Abromeit 1931-1940) has not been revealed in the present study. 1982


**FB34D** peatland near Miechzko [Milchbude] settl., forest south of Malinówka Wielka v. and east of Oraczew v., Elk comm. (Gross 1910, 1912; Abromeit et al. 1931-40; Źukowski 1969). Ex! († 2009) due to drainage and overgrowing with trees and reed. 1908 or 1909

**FB35B** fen bordering a small lake [Mechatssee] south-west of Czaple v., Elk comm. (Koppe and Koppe 1931). Ex! († 2009) due to eutrophication, drainage and overgrowing with reed. 1930

**FB35B** fen bordering the west part of Godle Lake near Chelchły [Chelchen] v., Czaple [Zappeln] v. and Przykopka [Birkenwald, Przykopken] v., Elk comm. (Koppe and Koppe 1931; Abromeit et al. 1931-40; Źukowski 1969). Ex! († 2009) due to overgrowing with trees and shrubs, drainage and peat extraction. The paper containing the original data of H. Steffen (see Abromeit et al. 1931-40) has not been revealed in the present study. 1936

**FB35C** fen bordering a small lake in a forest [Dallinitzwal] north-east of Elk town, Elk comm. (Gross 1910; Abromeit et al. 1931-40). Ex! († 2009) due to overgrowing with trees and drainage. 1908 or 1909


**FB45C** fen bordering Tatra Duże Lake [Grosse Tatarenaa] south of Elk, Elk comm. (Kalkreuth 1914; Abromeit 1928; Abromeit et al. 1931-40; Źukowski 1969). Ex! († 2009) due to overgrowing with trees and drainage. 1924

**FB50** fen bordering the east part of Beldany Lake, Ruciane-Nida comm. (KRA: leg. A. Kornaś; J. Kornaś 1952; Źukowski 1969). Ex! († 2009) due to overgrowing with trees. 1952

LITHUANIAN LAKE DISTRICT
FA77D fen adjacent to Grzybina v., Wiżajny comm. (! leg. 2008: 2). 2008**
FA78C fen adjacent to Stankuny v., Wiżajny comm. (! leg. 2008: 2). 2008
FA78C fen adjacent to Kolonia Wiżajny settl., Wiżajny comm. (! leg. 2008: 2). 2008
FA78C fen bordering the south-eastern part of Prudel Lake, Wiżajny comm. (! leg. 2008: 2). 2008
FA79C “Plinia” fen adjacent to Ejszerszysz v. and state border, Rutka-Tartak comm. (! leg. 2008: 2). 2008
FA88C poor fen bordering the north part of Czarne Lake by Smolniki v., Wiżajny comm. (BIL: leg. A.W. Sokołowski 1969, Sokołowski 1973). Ex? (! 2009), most probably due to overgrowing with trees and shrubs and peat extraction, but small patches of suitable habitat still exist. 1969
FA89D fen adjacent to Kociołki v., Szypliszki comm. (! leg. 2008: 3). 2008*
FA98D bog margin near Sidorówka v. near Jeleniwko, Jeleniewo comm. (Pawlikowski 2008a, ! leg. 2004: 1). 2004*


FB19D WPN: rich fen bordering the north-east part of Kruszyn (Krusznik) Lake, Nowinka comm. (Sokołowski 1990, Jutrzenka-Trzebiatowski et al. 2002). Ex? (! 2009) due to overgrowing with reed, trees and shrubs, but small patches of suitable habitat still exist. between 1992 and 1999

FB28D Rospuda river valley, rich fen west of the river, east of the Topoliówka v., Augustów comm. (! leg. 2006: 2). 2006

FB39A Rospuda river valley, forested mile east of the Rospuda river, south of the Blizna river and east of the Koźia Szyja hill, Augustów comm. (Sokołowski 1989, ! 2003, 2007, 2009: 3). This is possibly the place where W. Szafer collected the specimens in 1929 (KRA). 2009


FB39C rich fen adjacent to allotment gardens in Augustów, near “Szosa do Sejn” road, Augustów comm. (! leg. 2006: 2). 2006*


GA90C rich fen 1 km east of Nowe Boksze v., near Czarna river valley, Krasnopol comm. (! leg. 2004: 3). 2004**

GA90C margin of the bog south of Boksze Lake, Puńsk comm. (! leg. 2006: 1). 2006

GB00A fen south of Buczniel (Boczniel) Lake near Podlas v., Krasnopol comm. (! leg. 2004: 1). 2004**

GB00B poor fen bordering the north part of Płaskie Lake, Sejny comm. (WA: leg. L. Bogdanowicz 1975; Kłosowski and Tomaszewicz 1979; S. Kłosowski, pers. comm. in 2008). Ex? (! 2009) due to overgrowing with reed, trees and shrubs, but small patches of suitable habitat still exist. 1999

GB00B “Magdze Bagna” sparsely wooded rich fen bordering the south-west part of Płaskie Lake, Krasnopol comm. (! leg. 2004: 2). 2006*

GB00C rich fen between Długie Lake and a small lake northwards, Krasnopol comm. (! leg. 2003, ! 2006: 1). 2006*

GB01B fen bordering the western part of Dusajtis (Drafajtis) Lake (near Dusznica v.), Sejny comm. (! leg. 2004: 2). 2004


GB10D rich fen west of Sarnteki (Šarnteti) v., Giby comm. (! 2003-2007: 2). 2007*


GB11D rich fen 0.5 km south-west of the southernmost part of Wilkokuk Lake, Giby comm. (Sokolsky 1978, ! 2004: 2). 2004


GB12C rich fen “Mielubagno” bordering the north-east part of Zława Lake, Giby comm. (! leg. 2004: 2). 2004*


GB32A Augustowski Canal valley, fen adjacent to Rygol v., Płaska comm. (BIL: leg. J. Żurawski 1974). Ex (! 2009). There are several fens adjacent to that village and the precise location has not been revealed, but the species does not occur in any of them. 1974

NORTH POLSLIE [PODLACHIA] LOWLANDS

FB56D fen south of Belda v., Rajgród comm. (Żukowski 1969). Ex (! 2009) due to drainage and intensive meadow management. 1956

GB40 mere near Jastrząbna v., Sztabin comm. (Żukowski 1969). The locality is imprecise, but most probably Ex? (! 2002, 2009) due to drainage, intensive meadow management and forest development. 1954


LIST OF UNCERTAIN LOCALITIES

AD09 Jordanowo v., Świebodzin comm. (Żukowski 1969). The original paper of Ulbrich (1916) (see Żukowski 1969) does not contain information on the species and the area mentioned.

CC16 Smukala, present district of Bydgoszcz city, Bydgoszcz comm. (Zając and Zając 2001; ATPOL database after Bock 1908). The original paper of Bock (1908) does not contain information on the species.

DF35 Strzemieszycze, present district of Dąbrowa Górnicza city, Dąbrowa Górnicza comm. (Zając and Zając 2001; ATPOL database after herbarium specimens collected by A. Sendek in 1976). There is no other information on this locality.

FA84C peatland 2 km west of Goldap, Goldap comm. (Zukowski 1969). This locality is uncertain, as it can be identical with the listed above locality near Wilkajcie.

In Poland, the number of localities of the species recorded so far is estimated at 82. The main area of the species occurrence in Poland are the Lithuanian Lake District (50 localities), the eastern part of the Masurian Lake District (19 localities) and the Biebrza river valley in North Podlaskie Lowlands (6 localities). The distribution of the species in the area studied as well as the status of the localities are presented in Figure 1.

My survey revealed the presence of Baeothryon alpinum at 47 localities in the Polish Lowlands after the year 2000. Among them were 25 populations previously unknown, discovered by the author. Twenty-seven localities are definitively extinct, and 8 are most probably extinct. The main reasons for extinction are: overgrowing with trees and shrubs, drainage, peat extraction and expansion of reed.

There is clear disparity between the conditions of populations in the two neighbouring areas (Fig. 2). Considering the localities known before the year 2000 only, the Lithuanian Lake District has 14 (39 in total) populations confirmed after the year 2000 (56%), whereas in the Masurian Lake District only 3 populations have been confirmed (16%). Most of the localities in the Lithuanian Lake District are concentrated in the borderline areas of Augustów Forest, including the Sejny Lakeland and Wigry National Park, and in Góry Sudawskie region and adjacent areas.

Moderately abundant populations (several hundreds of shoots) predominate, comprising more than a half (57%) of the total number of populations. As many as 15% of the localities host very small populations (less than 100 shoots), but as many as 14 populations (nearly 30%) are numerous (more than 1000 shoots). They were noted in the Lithuanian Lake District (10 populations), as well as in the Masurian Lake District (2 populations) and the upper Biebrza area (2 populations). It is difficult to estimate the size of the biggest population (in the “Kobyla Biel” fen bordering Białe Augustowskie Lake), since it densely covers an area of more than 5 hectares. There are 28 localities (60%) that are threatened due to secondary succession, nine of them being severely threatened.

DISCUSSION

Young postglacial landscapes favour the development of peatlands (Tobolski 2003), which are the habitat of Baeothryon alpinum. Many of the mire plants in Poland occur mostly in lakeland areas in north and north-western part of the country (see Zając and Zając 2001). In the case of north-easternmost Poland, the high abundance of mire species is associated with greater availability of suitable habi-
tats (and postglacial history of plant cover – Środoň 1972). Moreover, the influence of boreal climate which is stronger in north-eastern Poland than in other parts of the country (Kondracki 2001) can favor the occurrence of boreal species, since many of the plants occurring in peatland ecosystems in Central Europe are bound primarily to the taiga zone (see Hultén and Fries 1986a, b). This is particularly true in the case of *B. alpinum* as nearly all the lowland localities of the species are situated in the lakeland areas of northern Poland, and are mostly aggregated in the north-easternmost part of the country.

The number of existing localities of *Baeothryon alpinum* known from the Lithuanian Lake District is surprisingly high in comparison with the adjacent Warmia and Masuria region, which shared similar geological (glacial) history and is geomorphologically similar (Kondracki 2001). Bloch-Orlowska (2007), who investigated *Carex chondorrhiza* (a mire species that shows a pattern of distribution dynamics similar to that of *B. alpinum*), pointed out that because of climate-related reasons the localities beyond the species main range are particularly vulnerable to accelerated decrease in number. Since many of the populations of *C. chondorrhiza* in north-eastern Poland still persisted, the author concluded that extinctions involved mostly the localities situated beyond the climatic optimum for the species. But if climatic factors explain why many mire species survived best in the north-easternmost Poland, the above pattern should be observed among boreal species only.

An argument against this theory is the example of *Liparis loeselii*, an important fen species not related to the boreal zone (it occurs mainly in Central Europe – Hultén and Fries 1986b). It has lost most of its previously recorded localities in Poland, and the majority of the populations survived in the Lithuanian Lake District in north-eastern Poland (Pawlikowski 2008b), where the boreal climatic influences are strongest. This suggests that independently of the species distribution, special attention should be given to man-related factors, particularly the intensification of land management (e.g. drainage). Despite different distribution patterns, both *B. alpinum* and *L. loeselii* have the most vital populations in north-easternmost Poland. Apart from the climatic factors, the most important reasons for this situation seem to be related to the fact that in the Lithuanian Lake District traditional extensive land management practices continued until very recently, whereas in many other lake-lands in north-eastern Poland (including Masuria and Warmia territory) the area was intensively managed and drained before and after World War II. This confirms the opinion of Jasnowski et al. (1968), who proved that the disappearance of mire plant species is caused primarily by human activity.

This pattern of decrease is true in the case of many other rare and endangered mire species with various distribution, such as *Saxifraga hirculus, Hammarbya paludosana, Malaxis monophylla, Betula humilis* and *Eriophorum gracile* (Kaźmierczakowa and Zarzycki 2001; Zając and Zając 2001). Furthermore, the explanation based on the regional differences in land-use intensity is strengthened by the fact that many species considered threatened in Poland, are not threatened at all in the Lithuanian Lake District (e.g. Sokolowski 1973, 1990; Kawecka 1991; Pawlikowski 2008a, c).

There have been little data on the population size of *Baeothryon alpinum* at the localities in Poland until now. Both very abundant (Tyszkowski 1992; Żukowski 2001; Bloch-Orlowska and Pisarek 2005) and small (Balcerkiewicz and Markowski 1969; Plackowski 1980; Dajdok and Kącki 1999; Pawlikowski 2008a) populations were recorded. The high number of shoots in the Lithuanian Lake District, with
at least 9 populations exceeding 1000 shoots, indicates that the Polish part of the Lithuanian Lake District hosts the majority of the Polish resources of the species. The estimated total number of shoots in the Lithuanian Lake District probably reaches 100 000, while in both neighbouring regions (Masurian Lake District, upper Biebrza valley) it does not exceed 10 000. The last two lowland populations confirmed during the past decades in southern and western part of Poland are on the verge of extinction.

In the part of Belarus adjacent to Poland, B. alpinum has not been recorded since the 19th century, and the species has the status of vulnerable species for the whole country (Skuratovich 2005). At the same time, it is considered not threatened at all in nearby Lithuania (Rašomavičius 2007). Thus, the aggregation of vital populations of Baeothryon alpinum in the Lithuanian Lake District with the adjacent areas is part of the species main range that includes nearby Lithuania (Hultén and Fries 1986a). This is in contrast to the remaining, scattered lowland localities beyond the main range of the species, all of them being extinct or near extinction.

As the number of populations of Baeothryon alpinum in Poland is decreasing, conservation measures should be undertaken where needed, and nature reserves established that would protect the most important (best preserved or most numerous) populations of the species. At present, apart from the six confirmed populations in the Biebrza National Park and the Wigry National Park, there are five confirmed populations situated within the borders of nature reserves. Among the localities that should be designated as nature reserves are:

1. fens in the lower Rospuda river valley;
2. “Kobyla Biel” fen bordering north-western part of Biale Augustowskie Lake;
3. “Bagno Parchacz” fen near the village of Stara Kamionka;
4. a fen north-east of the village of Kleszewo.

Populations (2-4), as well as those in the Biebrza National Park, need active protection (removing of shrubs and trees, sometimes also restoring proper hydrological conditions).

When the IUCN definitions of categories of threat (IUCN...2008) are applied, Baeothryon alpinum should be treated as Vulnerable (VU) in Poland, since A2c+3c criteria are met. When assessing the species status at a smaller geographical unit, the threat categories vary between the particular regions of lowland Poland. These are as follows:

- Critically Endangered (CR) – Masurian Lake District (criteria A2c+3c; B1ab(i-iv)+2ab(i-iv)) and Silesian Lowlands (criterion D; see Dajdok and Kački 2002);
- Vulnerable (VU) – Northern Podlasie Lowlands (criterion B1ab(i-iv)+2ab(i-iv)), but the category of threat is down-graded, since populations within the region may experience a “rescue effect” from populations in the neighbouring Lithuanian Lake District; see IUCN 2003);
- Near Threatened (NT) – Lithuanian Lake District (the species does not qualify for VU category now, but is close to qualifying for and is likely to qualify for a VU category in the near future).

In the South Pomerania Lake District Baeothryon alpinum has been considered critically endangered (Żukowski and Jackowiak 1995), but it has now become clear that the species should be regarded most probably extinct. Its only population in the “Trzęsacz” nature reserve has not been confirmed since 1983. It is beyond dispute that the species is extinct (Ex) in all the other regions of lowland Poland where it used to occur.

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LITERATURE CITED


